

Application of Modified Magnetite Nanoparticles as a New Sorbent for Separation/Preconcentration of Mercury(II) Trace Amounts and its Determination by Cold Vapor Atomic Absorption Spectrometry

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Abstract. A new, simple, fast and reliable method has been developed to separation/preconcentration of trace amounts of mercury ions using dithizone/sodium dodecyl sulfate-immobilized on alumina-coated magnetite nanoparticles (DTZ/SDS-ACMNP) and its determination by cold vapor atomic absorption spectrometry (CVAAS). This method avoided the time-consuming column-passing process of loading large volume samples in traditional solid phase extraction (SPE) through the rapid isolation of DTZ/SDS-ACMNP with an adscititious magnet. Under the optimal experimental conditions, the enrichment factor, detection limit, linear range and relative standard deviation (RSD) of Hg(II) ions were 250 (for 500 mL of sample solution), 0.058 µg mL⁻¹, 0.2–80.0 µg mL⁻¹ and 3.55 % (for 5.0 µg mL⁻¹, n = 10), respectively. The adsorbed mercury ions were quantitatively eluted by 2.0 mL of 0.4 mol L⁻¹ HBr solution. The presented procedure was successfully applied for determination of mercury content in the different samples of water and blood.

Keywords: Solid-phase extraction, alumina-coated magnetite nanoparticles, mercury(II), dithizone, cold vapor-atomic absorption spectrometry

INTRODUCTION

In recent years, there has been an increased concern over the concentration of mercury in drinking and natural water due to its high toxicity and pollution to the environment, and especially the aquatic system.^{1,2} Mercury is leached from rocks and soil into water system by natural processes, some of which are accelerated by human activities.

There are several analytical techniques for mercury determination at sub-ppb levels such as electroanalytical,³ inductively coupled plasma-atomic emission spectrometry (ICP-AES),⁴ inductively coupled plasma-mass spectrometry (ICP-MS),^{5,6} atomic fluorescence spectrometry⁷ and neutron activation analysis.⁸ Cold vapor-atomic absorption spectrometry (CV-AAS) is a very efficient, simple, low cost and widely used technique for accurate determination of sub-micrograms per milliliters of mercury.^{9–11} However, due to its low concentrations in numerous samples and high levels of non-toxic components that usually accompany analytes (especially marine samples), a cleanup and preconcentration step is often necessary prior to its measurement.¹²

The direct determination of trace metals especially toxic metal ions such as mercury, from various samples requires mostly an initial and efficient preconcentration step.¹³ This preconcentration is required to meet the detection limits as well as to determine the lower concentration levels of the analyte of interest.¹⁴ This can be performed simply in many ways including liquid and solid phase extraction (SPE) techniques.^{15,16} The application of SPE technique for preconcentration of trace metals from different samples results in several advantages such as the minimal waste generation, reduction of sample matrix effects as well as sorption of the target species on the solid surface in a more stable chemical form.¹⁷ An efficient SPE method should consist of a stable and insoluble porous matrix having suitable active groups (typically organic groups) that interact with heavy metal ions.¹⁸ So far, numerous substances have been applied for preconcentration of mercury, such as dithiocarbamate derivatives,^{19–21} cysteine,²² 4-(2-pyridylazo)-resorcinol,²³ dithioacetal derivatives,^{24,25} aliphatic amines 2-thiophenecarboxaldehyde,²⁶ etc. However, the main problems associated with these materials

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are time consuming and high cost for their preparation and operating difficulties. Exploring new functional materials are still necessary because of complexities of real samples and analysis demand.

At present, nanometer material has become more and more important due to its special properties. Although investigations of the surface chemistry of highly dispersed oxides, *e.g.* TiO₂, Al₂O₃, ZrO₂, SiO₂ and ZnO, indicated that these materials had very high adsorption capacity,^{27–31} but separation of these particles from aqueous medium is difficult because of very small dimension and high dispersion. In this study, a new method combining nanoparticles adsorption with magnetic separation has been developed and applied for the separation and preconcentration of mercury ions.^{32–35} Magnetite nanoparticles have been successfully applied to separate some proteins and organic compounds^{36,37} and some heavy metal ions.^{38–43} These methods were based on the solid-phase extraction of trace amounts of silver and lead ions using dithizone/sodium dodecyl sulfate-immobilized on alumina-coated magnetite nanoparticles.^{44,45} Surfactant molecules are known to form self-aggregates called “hemimicelles” or “admicelles” on solid alumina surfaces in acidic medium. Insoluble hydrophobic compounds can be solubilized in hydrocarbon cores of these micelles. In this work, sodium dodecyl sulfate (SDS) immobilized on alumina-coated magnetite nanoparticles was used for the collection of insoluble as a Chelating agent for preconcentration of Hg ions in water samples.⁴⁶

In this study, a new method combining nanoparticles adsorption with magnetic separation has been developed and applied for the separation and preconcentration of mercury ions. Dithizone (DTZ) was chosen as doping reagent because of its favorable coordination capacity and selectivity for Hg(II). It is concluded from the extraction data that the new sorbent is a promising material for the SPE of mercury. In this paper, we will explore the possibility of dithizone/sodium dodecyl sulfate immobilized on alumina-coated magnetite nanoparticles (DTZ/SDS-ACMNPs) to act as SPE sorbents for the separation/ preconcentration of trace level of mercury ions from environmental samples prior to determination by CV-AAS technique.

EXPERIMENTAL

Apparatus

A flame atomic absorption spectrophotometer (PG Instruments, model PG 990, England) equipped with a mercury hollow cathode lamp as light source and hydride vapor generator (WHG-103 A) was used for mercury generation and absorbance measurements. The mercury compounds were reduced to metallic mercury

Table 1. Applied conditions for mercury determination with the WHG-103 A (CV) system

Parameter	Applied conditions
Measurement mode	Continuous-flow method
Signal processing	Integrated hold
Integration time / s	10
Argon carrier gas flow rate ^(a)	80
Sample aspiration flow rate ^(a)	4.5
5.0 mol L ⁻¹ HCl flow rate ^(a)	2.1
0.5 % NaBH ₄ flow rate ^(a)	2.0

^(a) Expressed in mL min⁻¹.

with sodium tetrahydroborate and the mercury generator was operated with argon as carrier gas. All the measurements were carried out in the cold state under the conditions given in Table 1. A Fourier transform infrared spectrometer (FTIR Prestige-21, Shimadzu), scanning electron microscope (LEO 1455VP SEM) and vibrating sample magnetometer (VSM 7400 Model Lake-Shore) were used to characterize the structure of the prepared MNPs and DTZ/SDS-ACMNPs. The pH measurements were conducted by a Metrohm 780 pH-meter calibrated against two standard buffer solutions of pH = 4.0 and 9.0. Other instruments used were: mechanical stirrer (Heidolph, RZR2020), ultrasonic bath (S60H Elmasonic, Germany) and orbital shaker (Ika, KS130 Basic). In addition, for magnetic separations a strong neodymium-iron-boron (Nd₂Fe₁₂B) magnet (1.2 T, 2.5 × 5 × 10 cm) was used.

Reagents and Solutions

All of the chemicals were of analytical grade and all solutions were prepared by double distilled de-ionized water with electric conductivity below 1.2 μS cm⁻¹. All of the chemicals were obtained from Merck (Darmstadt, Germany). A stock 1000 mg mL⁻¹ of mercury (II) was prepared by dissolving 0.1354 g of HgCl₂ in 5 mL of concentrate nitric acid and was diluted to 100 mL. The pHs of the solutions were adjusted with phosphate buffer, ferrous chloride (FeCl₂·4H₂O), ferric chloride (FeCl₃·6H₂O), 1,5-diphenyl carbazole (DTZ), sodium dodecylsulfate (SDS), aluminum isopropoxide, ethanol, hydrobromic acid, ammonia, hydrochloric acid and sodium hydroxide were used without further purification processes.

Preparation of Alumina-coated Magnetite Nanoparticles (ACMNPs)

The Fe₃O₄ nanoparticles (MNPs) and alumina-coated magnetite nanoparticles (ACMNPs) were prepared by chemical coprecipitation methods as reported in our previous works.^{44,45} At first step the MNPs were

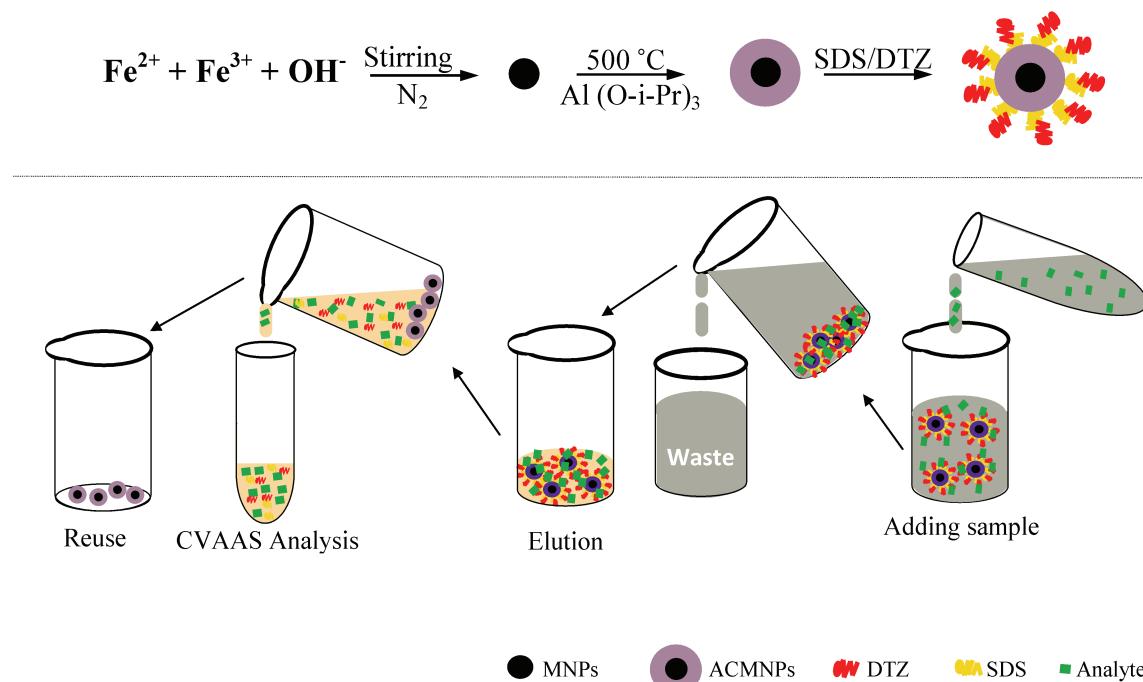


Figure 1. Procedure for magnetic solid-phase extraction.

prepared by chemical co-precipitation procedure. After the reaction, the obtained MNPs precipitate was separated from the reaction medium under the magnetic field, and rinsed with 200 mL pure deionized water four times. Then, the product was oven dried at 80 °C.

At the next step for preparation of ACMNPs, aluminum isopropoxide (1.0 g) was dissolved in ethanol (60 mL) to form a clear solution. MNPs (0.1 g) were then dispersed in the freshly prepared solution for 5 min with the aid of ultrasonic waves. A mixture of water and ethanol (volume ratio: 1:5) was added dropwise to the suspension of MNPs with vigorous stirring. The mixture was stirred for half an hour after the addition. Subsequently, the suspension was standing for one hour before separating and washing with ethanol. After five cycles of separation/washing/redispersion with ethanol, the powder obtained was oven dried and calcined at 500 °C for three hours.

General Procedure

The procedure for the magnetic extraction (MSPE) consists of three steps: firstly, 10 mL of DTZ/SDS solution (0.1/0.02 mmol L⁻¹) was added to 0.05 g ACMNPs and the pH was adjusted to about 2.0 by dropwise addition of 0.1 mol L⁻¹ HNO₃ solution, the mixture was shaken and for 2 min. Furthermore, the magnet was deposited at the bottom of the beaker and the DTZ/SDS coated ACMNPs were isolated from their suspension. Secondly, 10 mL of Hg(II) ion solution (10 µg mL⁻¹) was transferred to this sorbent and the pH value was adjusted to ~3.5 with phosphate buffer and, the soluti-

on was shaken for 5 min to facilitate adsorption of the metal ions onto the nanoparticles; and finally, the magnetic adsorbent was separated using magnet and the adsorbed ions were eluted with 2 mL HBr solution 0.4 mol L⁻¹. And this eluate solution was also separated for CV-AAS analysis. Illustration of the whole procedure of the preparation of ACMNPs and its application as MSPE sorbents for enriching analyte can be followed in Figure 1.

Sample Preparation Procedure for Water

Water samples (*i.e.*, tap water, river and spring water), were filtered through filter paper (Whatman, No. 4) to remove suspended particulate matter after collection and acidified to a pH of about 1.0 with concentrated HNO₃ prior to storage in polyethylene container for use. Subsequently were neutralized with concentrated NH₃ and then pH of solutions were adjusted to 3.0. The SPE procedure was carried out as described in general procedure

Sample preparation procedure for blood

Five samples of blood were obtained from healthy human bodies by in the sterile blood collecting tubes containing EDTA as anticoagulant. 4 mL conc. HNO₃ was added to 5 mL from each samples and heated for 15 minutes at 80°C, and then 2 mL H₂O₂ was added and heated for 20 minutes again. After filtering and eluting with 10 mL diluted nitric acid, the resultant solution was neutralized and diluted. Then pH was adjusted to 3.0, and it was analyzed according to the general procedure for extraction and determination of mercury content.

RESULTS AND DISCUSSION

Characterization of MNPs, ACMNPs and DTZ/SDS-ACMNPs

It is most important that the ACMNPs as sorbent should possess suitable magnetic property. This property was studied by measuring the hysteresis and remanence curve by means of a vibrating sample magnetometer (VSM). The curve shows that ACMNPs with value of large saturation magnetization of 9.34 emu g^{-1} exhibit typical superparamagnetic behavior due to no hysteresis (Figure 2). Therefore, sorbent of ACMNPs is sufficient for magnetic separation with a conventional magnet.

A scanning electron microscope (SEM) was used to examine the surface of the MNPs and ACMNPs as adsorbent (Figure 3). The image of Figure 3b illustrates a highly porous morphology of adsorbent with uniform size distribution of the nanospheres.

The modified ACMNPs were confirmed by FT-IR analysis, as shown in Figure 4. In the spectrum of ACMNPs (Figure 4b), compared with the spectrum of MNPs, after binding alumina, and the broadening of the peak at 638.07 cm^{-1} can be assigned to Al–O, that overlapped with Fe–O characteristic peak.⁴⁷ Comparison of the FT-IR spectra of ACMNPs and DTZ/SDS-ACMNPs (Figure 4c) is also shows a new sharp peak at 1375.96 cm^{-1} appeared, it was due to that the C–N stretching peak of DTZ stabilized on ACMNPs. Consequently, the FT-IR data suggest that DTZ are successfully immobilized on the ACMNPs surface.

Amounts of SDS and DTZ

Hiradie *et al.* proposed that water-insoluble chelating sorbents could be trapped into the aggregate of sodium dodecyl sulfate (SDS) on alumina particles.⁴⁸ The molecules of anionic surfactants can effectively be sorbed on the positively charged surface. ACMNPs have positively charged surfaces in highly acidic solutions that can strongly adsorb a negatively charged surfactant such as

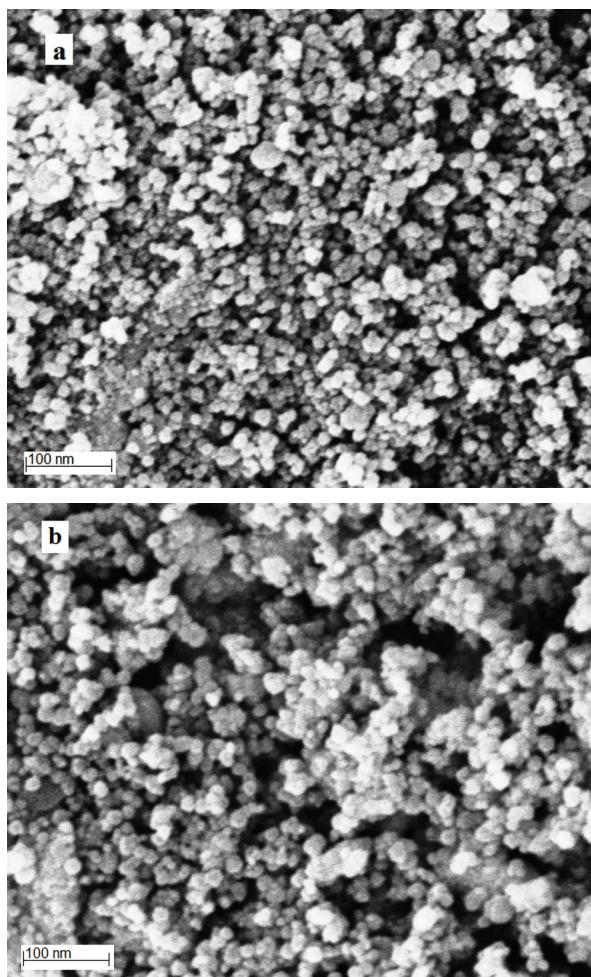


Figure 3. SEM images of MNPs (a) and ACMNPs (b).

SDS. A concentration of SDS, below its critical micellar concentration (CMC, 8 mmol L^{-1}), was used. Above this concentration, the excess of SDS would form micelles in the aqueous solution, which were not adsorbed on alumina surfaces. Results show that with the increase of SDS concentration, the absorbance increases and a maximum is obtained after the SDS concentration approaches

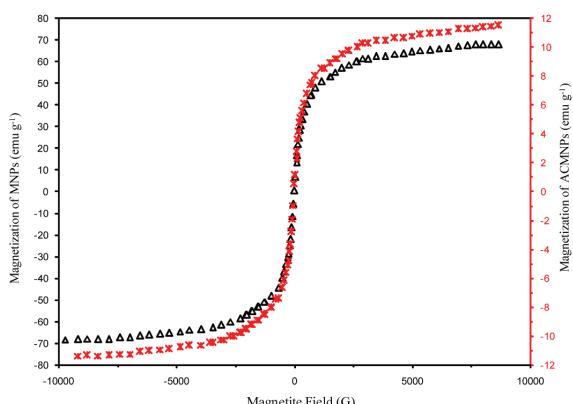


Figure 2. The magnetic behavior of MNPs (Δ) and ACMNPs ($*$).

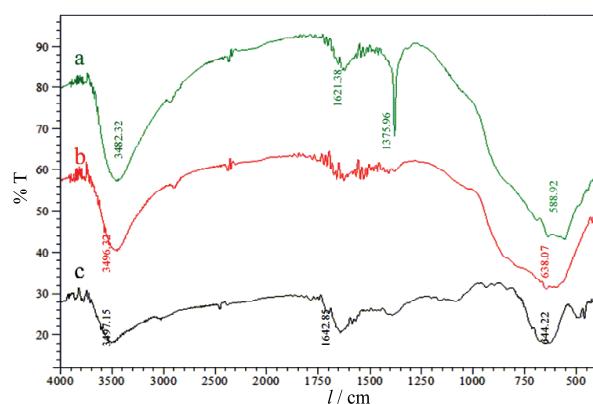


Figure 4. FTIR spectra of the MNPs (a), ACMNPs (b), and DTZ/SDS-ACMNPs (c).

to 6.0×10^{-2} mmol L⁻¹ and remains constant up to CMC and then decreased, because above this point micelles are formed strongly. Therefore, 0.1 mmol L⁻¹ SDS concentration was employed for further experiments.

In order to study the effect of DTZ concentration on adsorption of mercury ions on the ACMNPs, ammoniacal solutions of DTZ/SDS with constant concentration of SDS and different concentrations of DTZ were used. At DTZ concentrations less than 2.0×10^{-2} mmol L⁻¹, the amount of DTZ molecules immobilized into SDS cores are too low to completely complex all mercury ions, so recoveries less than 100% were observed. At concentrations more than 2.0×10^{-2} mmol L⁻¹ of DTZ, the sorbent sites are too rich, with respect to DTZ molecules, to allow mercury ions to be adsorbed by formation of its DTZ complex. Therefore, 2.0×10^{-2} mmol L⁻¹ of DTZ was selected as the optimum concentration for further studies.

Effect of pH

The effect of pH was studied for the recovery of 10 mL mercury ion ($10 \mu\text{g mL}^{-1}$) on 0.05 g of DTZ/SDS-ACMNPs (Figure 5). A series of Hg(II) solutions with different pH values from 2.0 to 12.0 were used, and the adsorbed mercury was eluted with 2 mL of 0.4 mol L⁻¹ HBr solution. Results show that the adsorption behavior of metal ions was sensitive to pH changes. Proton in acid solution can protonate binding sites of the chelating molecules, and hydroxide in basic solution may complex and precipitate metal ions.⁴⁹ Moreover, it is evident that the adsorption of mercury is quantitative (> 90%) at a pH range of 2.5–7.0. Therefore, a pH of 3.5 was selected as the optimum pH.

Adsorption Isotherm

The equilibrium isotherm of mercury ions adsorption by the SDS/DTZ-ACMNPs in acetate buffer solution at pH 3.5 and 25 °C is shown in Figure 6a. The adsorption behavior could be described by the Langmuir adsorption

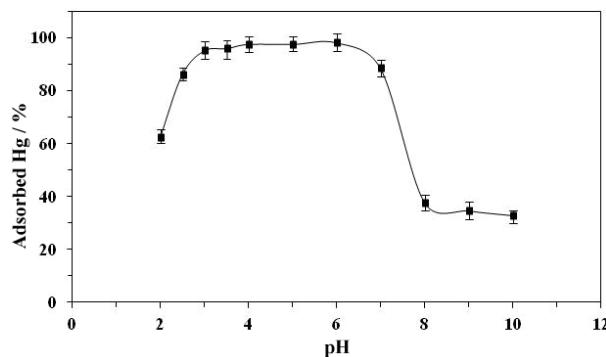


Figure 5. The effect of pH on the Hg(II) adsorption on SDS/DTZ-ACMNPs. Conditions: SDS/DTZ-ACMNPs (0.05 g), Hg(II) solution (10 mL, $10 \mu\text{g mL}^{-1}$), temperature: 25 °C.

equation:

$$\frac{C}{Q} = \frac{1}{Kq_m} + \frac{C}{q_m} \quad (1)$$

where Q is the equilibrium adsorption amount of Hg(II) (mg g⁻¹), C is the equilibrium mercury ion concentration in the solution ($\mu\text{g L}^{-1}$), q_m is the maximum adsorption amount of Hg(II) per gram of adsorbent (mg g⁻¹) and K is the Langmuir adsorption equilibrium constant (mL ng⁻¹). A plot of C/Q vs. C yielded a straight line (Figure 6b). From the slope and intercept of the line, the values of q_m and K can be estimated to be 21.27 mg g⁻¹ and 0.94 mL ng⁻¹, respectively.

Standing and Magnetic Separation Time

We found that the standing time had obvious effect on the SPE of target analyte. When the standing time were adjusted to 1, 3 and 5 min, recoveries improved to 81, 90 and 98%; respectively. While, if the ACMNPs are isolated immediately without a standing process, the recovery of Hg(II) ions is only 40%. Therefore, 5 min was sufficient to achieve satisfactory adsorption and better recovery of mercury ion. Meanwhile, in the

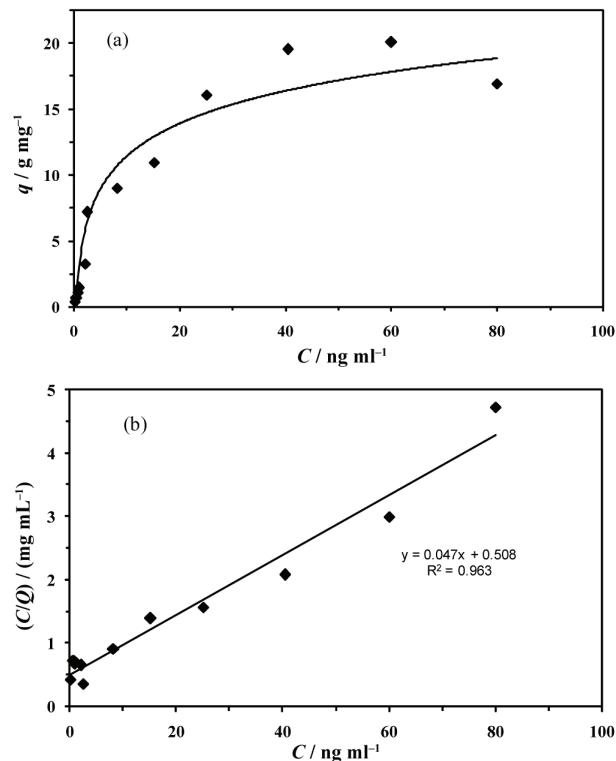


Figure 6. Adsorption isotherm of mercury on SDS/DTZ-ACMNPs (a) Plot of C/Q against C for the adsorption of mercury on SDS/DTZ-ACMNPs. Conditions: ACMNPs (0.05 g), DTZ/SDS solution (10 mL, 3×10^{-2} mmol L⁻¹ / 0.1 mmol L⁻¹, pH = 3.0), Hg(II) solution (50 mL, 0.2 – $80 \mu\text{g mL}^{-1}$); equilibrium time: 10 h, temperature: 25 °C.

experiment, DTZ/SDS-ACMNPs possessed superparamagnetism properties and large saturation magnetization, which enabled them to be completely isolated at the least time (less than 1 min) by a strong magnet.

Elution Condition, Maximum Sample Volume and Enrichment Factor

A satisfactory eluent should effectively elute the sorbed analytes with small volume, which is needed for a high enrichment factor, and should not affect the accurate determination of the analytes and destroy life time and reusability of solid phase. Since the protonation of the amine groups under strong acid conditions, the coordination interaction of chelated Hg(II) ions could be easily disrupted and subsequently Hg(II) ions were released from loaded sorbents into desorption medium. So various concentrations (0.1–1.0 mol L⁻¹) and volumes (0.5–10 mL) of HBr, HNO₃ and HCl solutions were used for the desorption of retained Hg(II). The results show that maximum recoveries obtained in the optimum concentration of these solutions were 98.5, 84.7 and 87.0 % for HBr, HNO₃ and HCl, respectively. Results are shown, at a concentration more than 0.4 mol L⁻¹ HBr, mercury ions were completely desorbed from sorbent surfaces. Therefore, 2.0 mL of 0.4 mol L⁻¹ are selected as the optimum volume and concentration of HBr, respectively.

Due to low concentration of heavy metals, preconcentrations have been done on large volume of real samples. In order to explore the possibility of concentrating Hg (II) from low concentration of analytes in large volumes of solution and acquire a high enrichment factor, investigating the effects of sample volume on the retention of mercury ions is of great necessity. According to the proposed procedure, various volumes in the range of 50–1000 mL of sample solutions were tested (Figure 7) in which the total amount of Hg(II) ions was constant at 5.0 µg. The maximum sample volume can be up to 500 mL with the recovery > 95 %. The further increase of the sample volume led to significant decrease of the recovery. Therefore, 500 mL of the sample solution was adopted as the maximum sample volume for the preconcentration of Hg(II) from sample solutions. Thus, the higher enrichment factor was 250 since the final elution volume was 2.0 mL.

Effect of Co-existing Ions

The optimal experimental conditions described above were used to study whether other co-existing ions that could act as interferents during the separation/preconcentration and analyte determination steps of the two-step method. The recovery of 10.0 µg L⁻¹ of mercury ions was investigated in binary mixtures containing mercury ion and one of the foreign ions. The following excess of ions did not interfere the reaction (*i.e.*, caused

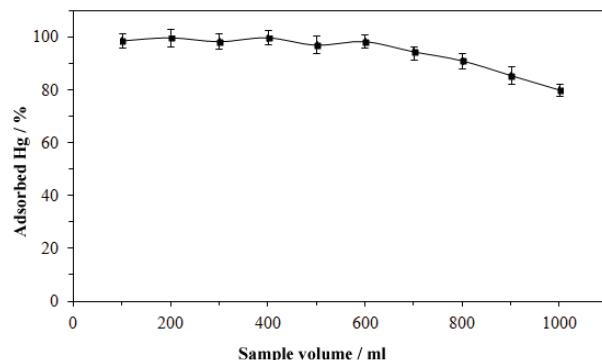


Figure 7. Effect of the sample volume on the recovery of Hg(II) from the SDS/DTZ-ACMNPs. Conditions: SDS/DTZ-ACMNPs (0.05 g), Hg(II) solution (0.1 µg, pH = 3.5), HBr as eluent solution (0.4 mol L⁻¹), temperature: 25 °C.

a relative error of less than 5 %): more than a 10000-fold amount of Na⁺, Ca²⁺, Mg²⁺ and K⁺; a 1000-fold amount of Mn²⁺, Ni²⁺, Zn²⁺, Cu²⁺, Fe²⁺, Co²⁺, Cd²⁺, K⁺, Cr³⁺, Fe³⁺, Bi³⁺, Pd²⁺ and NH₄⁺; a 500-fold amount of, NO₃⁻ and CH₃COO⁻, a 200-fold amount of SO₄²⁻, PO₄³⁻, C₂O₄²⁻, Ag⁺, Pb²⁺ and a 50-fold amount of I⁻, F⁻, Cl⁻ and Br⁻. The results showed that most of the investigated ions do not interfere in the adsorption–desorption and determination of traces of mercury ion in water samples.

Reproducibility in Preparation of DTZ/SDS-ACMNPs and Sorbent Regeneration

The reproducibility in preparation of DTZ/SDS-ACMNPs at same conditions for immobilized NPs that were produced different times was researched (constant in other conditions) and its effect at recovery was investigated and a relative standard deviation (RSD) < 6 % observed in results.

Regeneration is one of the key factors for evaluating the performance of the sorbents. In this work, it was found that the ACMNPs can be reused up to five times without loss of analytical performance. This reusable number is suitable because 4.0 g of ACMNPs could be prepared in one batch and only 0.05 g of ACMNPs was used for one extraction operation.

Analytical Performance and Method Validation

In order to show the validation of the proposed method, under the optimal experimental conditions, the analytical features of the method such as limit of detection (LOD), linear range of the calibration curve and precision were examined. The LOD of the proposed method based on three times the standard deviation of the blank (3S_b), was 0.058 µg mL⁻¹ for mercury ion (*n* = 10). The linear range of calibration curve for Hg(II) was 0.2–80.0 ng mL⁻¹ with a correlation coefficient of 0.9965. The regression equation for the line was $A = 0.093 C_{\text{Hg}} + (5.6 \times 10^{-3})$ (*n* = 5), where C_{Hg} is the concentration of

Table 2. Recoveries results of real water samples spiked with mercury ions

Sample	$\gamma(\text{Hg(II)}) / \mu\text{g L}^{-1}$		Recovery / %
	Added	Found	
Tap water (Sirjan)	0	—	—
	10	9.7 (± 0.2)	97.4
	20	20.2 (± 0.4)	101.1
River water (Hajibabad, Bandar Abbas)	0	—	—
	10	9.9 (± 0.3)	99.0
	20	20.4 (± 0.3)	102.5
Spring water (Koran, Sirjan)	0	—	—
	10	9.8 (± 0.1)	98.3
	20	20.1 (± 0.2)	100.5

Table 3. Mercury contents in various blood samples

Sample	$\gamma(\text{Hg(II)}) / \mu\text{g L}^{-1}$		Recovery / %
	Added	Found	
1	0	3.4 (± 0.4)	—
	5	8.6 (± 0.3)	97.8
2	0	7.5 (± 0.6)	—
	5	12.3 (± 0.5)	96.2
3	0	4.3 (± 0.7)	—
	5	9.4 (± 0.2)	102.4
4	0	2.7 (± 0.3)	—
	5	8.0 (± 0.4)	106.4
5	0	8.3 (± 0.5)	—
	5	12.2 (± 0.6)	98.4

Table 4. Comparison of the proposed method with other reported methods for separation/preconcentration of mercury ions

Adsorbent	Detection method	Enrichment factor	RSD / %	Linear range / $\mu\text{g mL}^{-1}$	Detection limit / $\mu\text{g mL}^{-1}$	Ref.
2,6-Pyridine dicarboxylic acid/nanometer silica	ICP-AES	175	3.0	—	0.09	(50)
Dithizone/neutral alumina	Spectrophotometry	100	2.5	—	4.0	(51)
2-Mercaptobenzimidazole/agar	CV-AAS	100	2.6	0.04–2.4	0.02	(52)
Dithizone/neutral alumina	CV-AAS	100	1.7	0.6–5.0	0.11	(53)
<i>N</i> -(2-chlorobenzoyl)- <i>N</i> -phenylthiourea loaded/sulfur	CV-AAS	333	3.9	0.02–1.20	0.012	(54)
Dithizone/cellulose	Spectrophotometry	33	3.5	0–2	2.0	(55)
Dithizone/silica gel	Spectrophotometry	500	3.0	1.0–1500	0.9	(56)
1,5-Bis(2-pyridyl)-3-sulphophenyl methylene thiocarbonohydrazone/Resin	CV-AAS CV-ETAAS ^(a)	28 42	3.4 3.5	0.1–30.0 0.009–1.5	0.01 0.006	(57)
1,4-Bis(4-pyridyl)-2,3-diaza-1,3-butadiene/Octadecyl silica cartridge	CV-AAS	128	2.4	0.13–25.0	0.05	(58)
Dithizone/ACMNPs	CV-AAS	250	3.55	0.2–80.0	0.058	This work

^(a) Cold vapor-electrothermal atomic absorption spectrometry

Hg(II) in $\mu\text{g mL}^{-1}$ and A is the absorbance. The RSD for 10 replicate measurements of $5.0 \mu\text{g mL}^{-1}$ of mercury ion was 3.55 %.

Analytical Applications

In order to check the applicability of the proposed method it was applied to the separation/preconcentration and determination of mercury in different samples of water and human's blood samples. The results tabulated in Tables 2 and 3. According this data, the added mercury ions can be quantitatively recovered from these samples by the proposed procedure. These results demonstrate, the presented procedure can be reliably applied for the separation/preconcentration and determination of Hg(II) in different samples of water and blood.

CONCLUSION

In this study, the new adsorbent (DTZ/SDS-ACMNPs) was prepared easily and low-costly utilized conveniently and harmless to environment. This sorbent was successfully applied for convenient, fast, simple and efficient enrichment of trace amounts of mercury ion from environmental water and blood samples. Magnetic separation in the method shortened analysis times greatly. The main benefits of this methodology are: no use of toxic organic solvent, simplicity and low cost. Furthermore, it avoids the time-consuming column passing (about 1 h in conventional SPE method) and filtration operation, and no clean-up steps were required. Table 4 shows a comparison of the proposed method with other reported SPE methods. It could be seen that some obtained values for

the proposed method such as linear range, LOD, sorbent capacity and enrichment factor are as or better than some of the previously reported methods. Moreover, total time of analysis of this method is reduced.

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