# Levels of Nanosize Radon Decay Products in Indoor Air: A Comparison for Different Environments

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# ABSTRACT

The fraction of unattached  $(f_{un})$  nanosized <sup>222</sup>Rn short-lived decay products in indoor air has been measured in different environments and the following ranges obtained: 0.03–0.25 in kindergartens, 0.03–0.18 in schools, 0.08–0.20 in wineries, and 0.12–0.68 in a karst cave. The dependence of  $f_{un}$  on the environmental conditions (air temperature, relative humidity and pressure), as well as on the working regime has been studied and will be discussed. Applying the dosimetric approach, dose conversion factors were calculated, based on the measured  $f_{un}$  values, and compared to the values, obtained from epidemiological studies and currently recommended to be used in radon dosimetry.

Key words: radon, decay products, unattached fraction, indoor air, dose conversion factors

## Introduction

The radioactive noble gas radon (<sup>222</sup>Rn isotope,  $\alpha$  decay,  $t_{1/2} = 3.82$  days) is always accompanied by its short--lived decay products (RnDP): <sup>218</sup>Po ( $\alpha$  decay,  $t_{1/2} = 3.10$ min),  $^{214}\text{Pb}$  ( $\beta/\gamma$  decay,  $t_{1/2}$  = 26.8 min),  $^{214}\text{Bi}$  ( $\beta/\gamma$  decay,  $t_{1/2}$ = 19.9 min), and  $^{214}\text{Po}$  ( $\alpha$  decay,  $t_{1/2}$  = 164  $\mu\text{s}).$  Initially, these products are positively charged free ions which after neutralization appear in air as nanosize clusters, the so-called unattached RnDP (in the size range 0.5–3 nm), which sooner or later, depending on the environmental conditions (air temperature, humidity, concentration and size distribution of aerosols), attach to aerosols and form attached RnDP in the size range between 200 and 800 nm<sup>1</sup>. Because of plate-out of aerosols on the walls and floor of a room, as well as air movement and entry of fresh air, radioactive equilibrium between RnDP and Rn in indoor air is only partly achieved and is expressed as a fraction between 0 and 1, called the equilibrium factor<sup>2</sup>, F.

It has become well known that, on the world wide average, breathing air with radon short-lived decay products contributes more than half to the effective dose a member of the general public receives from all natural radioactive sources of ionizing radiation, and that they are a major cause of lung cancer, second only to cigarette smoking<sup>3</sup>. To quantify the detrimental effects of Rn and/or RnDP to the human health, the so called dose conversion factor (DCF) is required, which relates the exposure to RnDP (expressed either in WLM or Bq m<sup>-3</sup> h) and effective dose (expressed in mSv). The old but still widely used unit, 1 WLM (working-level-month) is the exposure resulting from 170 hours breathing air with an activity concentration of short-lived radon decay products of 1 WL (working-level). 1 WL<sup>2</sup> was originally defined as the activity concentrations of RnDP which are in radioactive equilibrium (F = 1) with 100 pCi L<sup>-1</sup> (3700 Bq m<sup>-3</sup>) of <sup>222</sup>Rn, resulting in a potential alpha energy concentration of  $1.3 \times 10^8$  MeV m<sup>-3</sup>. DCF values have been obtained from epidemiologic studies on uranium miners. At present, the International Commission for Radiological Protection (ICRP) in Publication 65<sup>4</sup> recommends 5 mSv WLM<sup>-1</sup> for working and 4 mSv WLM<sup>-1</sup> for living environments. The epidemiology-based DCF will be hereafter in the text denoted by  $DCF_{\rm E}$ .

Birchall and James<sup>5</sup> elaborated a dosimetric approach to calculate the dose conversion factor based on  $f_{un}$ ; it will be hereafter in the text denoted by  $DCF_D$ . Their 'best estimate' of  $DCF_D$  for the indoor air conditions in dwellings was 15 mSv WLM<sup>-1</sup>. It thus appears that  $DCF_D$  is 3-fold

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higher than  $DCF_{\rm E} = 5 \text{ mSv WLM}^{-1}$ . The discrepancy between  $DCF_{\rm D}$  and  $DCF_{\rm E}$  values has not been fully clarified, the most probable reason originating in too high a value being chosen for the radiation weighting factor<sup>6</sup>,  $w_a$ . In addition, Porstendörfer<sup>7</sup> has shown that  $DCF_{\rm D}$  can be calculated separately for mouth  $(DCF_{\rm Dm})$  and nasal  $(DCF_{\rm Dm})$  breathing, using empirical equations:

$$DCF_{\rm Dm} = 101 \times f_{\rm un} + 6.7 \times (1 - f_{\rm un})$$
 (1)

$$DCF_{\rm Dn} = 23 \times f_{\rm un} + 6.2 \times (1 - f_{\rm un})$$
 (2)

In this paper, the results of our studies on  $f_{\rm un}$  in 29 rooms of kindergartens, 26 rooms of elementary and high schools, 4 rooms in wineries and at two points in the Postojna Cave are reported.  $DCF_{\rm D}$  values, calculated based on the Porstendörfer formulae (Eqs. 1 and 2), are discussed and compared with the  $DCF_{\rm E}$  value

recommended by ICRP-65.

#### **Materials and Methods**

### Measurement sites

Radon was surveyed in the indoor air of practically all the kindergartens (730) and schools (890) in Slovenia, within the national radon program, between 1990 and 1994<sup>8,9</sup>. For the study of  $f_{un}$ , those kindergartens and schools were selected in which elevated radon levels had been observed or were expected, based on geology<sup>10</sup>. Measurements, lasting for 1–2 weeks at each place, were carried out in 29 rooms of 13 kindergartens (in the period from March 1998 to February 2001) and in 26 rooms of 16 schools (in the period from March 1998 to April 2001).

Because of the elevated radon levels, permanent radon monitoring at the lowest point and the railway station in the Postojna Cave was introduced in 1995<sup>11</sup>. Based on the radon concentrations measured, working time for the employees in the cave is limited in order to keep their effective doses acceptably low. Measurements of  $f_{un}$ , lasting for 1–3 weeks were carried out at two places, at the lowest point (LP) of the cave in summer (when radon levels are highest: August 10–18, 1998; June 30 – July 8, 1999; July 19 – August 3, 2001) and in winter (when radon levels are lowest: December 14–22, 1998; December 10–20, 1999), and at the railway station (RS) in summer (June 3–18, 2001). In the context of the radon survey in underground premises of eight major Slovene wineries,  $f_{\rm un}$  was monitored for 1–2 weeks in the period from April to July, 2002.

#### Measuring technique

Portable SARAD EQF3020 and EQF3020-2 devices (SARAD, Dresden, Germany) were used. Air is pumped for 6 min at a flow rate of 2.4 dm<sup>3</sup> min<sup>-1</sup> over a metal mesh grid where unattached and attached RnDP are separated and then deposited electrostatically on two separate 150 mm<sup>2</sup> semiconductor detectors. Applying the Markov method<sup>12</sup>, the device gives Rn concentration in Bq m<sup>-3</sup> ( $C_{\rm Bn}$ ), and, separately for the unattached and attached form of RnDP, the following: individual concentrations in Bq m<sup>-3</sup> ( $C_{218Po}$ ,  $C_{214Pb}$ ,  $C_{214Bi} = C_{214Po}$ ), equilibrium equivalent activity concentration in Bq  $m^{-3}$  (defined as:  $EEC_{RnDP} = C_{RnDP} = 0.1065C_{218Po} + 0.515C_{214Pb} +$  $0.379C_{214Bi}$ ), potential alpha energy concentration in MeV m<sup>-3</sup> (defined as:  $PAEC = 3690C_{218Po} + 17830C_{214Pb}$ +  $13120C_{214Bi}$ ), *F* (defined as *F* = *PAEC* /(34640 C<sub>Rn</sub>), and  $f_{\rm un}$  (defined as  $f_{\rm un} = PAEC_{\rm un}/(PAEC_{\rm un} + PAEC_{\rm att})$ , with subscripts 'un'– unattached and 'att' – attached)<sup>13</sup>.

The data was transferred to a personal computer in the laboratory for evaluation. Origin 6.1 Data Analysis and Graphing Software was used for statistical data evaluation and presentation.

## Results

Diurnal variations of  $C_{\text{Rn}}$ ,  $C_{\text{RnDB}}$  F and  $f_{\text{un}}$  for a selected school are presented in Figure 1. Values of  $C_{\text{Rn}}$ ,  $C_{\text{RnDP}}$  are highest during the night, decrease during a working day and gain higher values again during the night. Values are kept constantly higher during weekends, e. g., from 14.4. to 16.4. High values of F are accompanied by low values of  $f_{\text{un}}$ . Similar situations were observed in all kindergartens and schools, although the absolute values of the measured parameters were different and so were the amplitudes of daily fluctuations. Table 1 shows arithmetic mean (AM) and arithmetic standard deviation values (ASD) of the measured parameters, calculated for the whole period of measurements.

The arithmetic mean values of the measured parameters resulting from continuous measurements in the

TABLE 1

INDOOR AIR RADON CONCENTRATION ( $C_{Rn}$ ), EQUILIBRIUM FACTOR BETWEEN RADON AND RADON DECAY PRODUCTS (F) AND UNATTACHED FRACTION OF RADON DECAY PRODUCTS ( $f_{un}$ ) IN SLOVENIAN SCHOOLS AND KINDERGARTENS, WITH THE NUMBER OF VALUES (n), ARITHMETIC MEAN (AM), ARITHMETIC STANDARD DEVIATION (ASD), *t*-VALUE (*t*) AND p-VALUE (*p*) OF THE *t*-TEST

		n	AM	ASD	t	р	
$C_{ m Rn}$	schools	26	1362	1718	2.372	0.021	significantly
	kindergartens	29	584	394			different
F	schools	26	0.61	0.15	2.024	0.048	significantly
	kindergartens	29	0.53	0.15			different
$f_{ m un}$	schools	26	0.13	0.03	2.114	0.039	significantly
	kindergartens	29	0.15	0.06			different

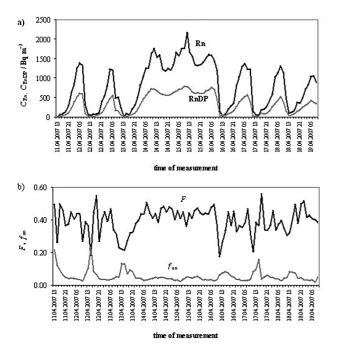


Fig. 1. Results of a continuous monitoring in a selected school, showing diurnal variations of concentrations of radon  $(Rn, C_{Rn})$ and radon short-lived decay products  $(RnDP, C_{RnDP})$ , equilibrium factor between Rn and RnDP (F) and unattached fraction of radon decay products  $(f_{un})$ .

Postojna Cave are listed in Table 2.  $C_{\rm Rn}$  values were higher in summer than in winter at the lowest point, and higher at the lowest point than at the railway station. On the other hand, F was lower in summer than in winter at the lowest point and lower at the lowest point than at the railway station in summer.

Results of continuous measurements in a selected winery are shown in Figure 2. In contrast to kindergartens and schools, no typical diurnal fluctuations have been observed in wineries. Arithmetic mean values of  $C_{\rm Rn}$ , F and  $f_{\rm un}$  are collected in Table 3.

#### Discussion

Figure 3 shows that  $C_{\text{Rn}}$  values both in kindergartens and schools fit well lognormal distribution. Table 1 dis-

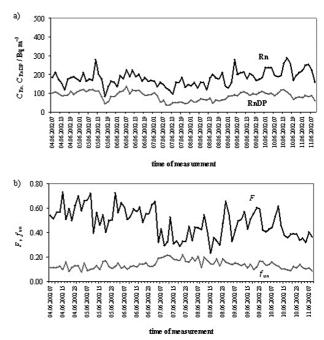


Fig. 2. Results of a continuous monitoring in a selected winery, showing diurnal variations of concentrations of radon  $(Rn, C_{Rn})$ and radon short-lived decay products  $(RnDP, C_{RnDP})$ , equilibrium factor between Rn and RnDP (F) and unattached fraction of radon decay products  $(f_{un})$ .

plays results of the *t*-test comparison of the arithmetic mean values of  $C_{\rm Rn}$ , *F*, and  $f_{\rm un}$  for the whole period of measurement. Radon levels are significantly higher in schools than in kindergartens. The probable reason lies in the age of buildings. While the majority of kindergartens were built after 1980, the school buildings are generally older. Here, due to aging, the basic concrete slab has numerous cracks and fissures and is no longer a sufficient barrier to radon. Further, there is very often no concrete slab at all in old buildings.

The difference in F levels in schools and in kindergartens most probably reflects the different working regimes at the two places. While in a classroom in a school the movement of students is minimal during class hours, children move more freely in the play room of a kindergarten, causing air movement and hence increasing pla-

TABLE	2
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THE POSTOJNA CAVE: AVERAGE VALUES (ARITHMETIC MEAN AND ARITHMETIC STANDARD DEVIATION) OF RADON CONCENTRATION ( $C_{\rm Rn}$ ), EQUILIBRIUM FACTOR BETWEEN RADON AND RADON DECAY PRODUCTS (F) AND UNATTACHED FRACTION OF RADON DECAY PRODUCTS ( $f_{\rm un}$ ), MEASURED DURING DIFFERENT PERIODS AT THE LOWEST POINT (LP) AND RAILWAY STATION (RS)

Season, year (site)	$C_{ m Rn}$ / Bq m <sup>-3</sup>	F	$f_{un}$
Summer, 1998 (LP)	$4090\pm440$	$0.34\pm0.06$	$0.58\pm0.14$
Winter, 1998 (LP)	$1470\pm1060$	$0.56\pm0.10$	$0.10\pm0.07$
Summer, 1999 (LP)	$4540\pm 600$	$0.35\pm0.09$	$0.60\pm0.16$
Winter, 1999 (LP)	$2070 \pm 1160$	$0.58\pm0.11$	$0.14\pm0.08$
Summer, 2001 (LP)	$5900\pm 620$	$0.35\pm0.07$	$0.68\pm0.11$
Summer, 2001 (RS)	$3330\pm370$	$0.63\pm0.13$	$0.15\pm0.05$

 TABLE 3

 AVERAGE VALUES OF RADON CONCENTRATION (C<sub>Rn</sub>), EQUILIBRIUM FACTOR BETWEEN RADON AND RADON DECAY PRODUCTS (F)

 AND UNATTACHED FRACTION OF RADON DECAY PRODUCTS (f<sub>un</sub>) OBTAINED IN SLOVENIAN WINERIES, WITH THE RESULTING DOSE CONVERSION FACTORS FOR NASAL BREATHING (DCF<sub>Dn</sub>)

Winery	Date in 2002	$C_{ m Rn}$ / Bq m <sup>-3</sup>	F	$f_{ m un}$	$DCF_{\rm Dn}$ / mSv WLM <sup>-1</sup>	$DCF_{ m Dn}/5$
01-S-05	23.04 13.05.	$91\pm20$	$0.25\pm0.08$	$0.20\pm0.06$	9.56	1.91
04-P-02	30.0514.06.	$182\pm37$	$0.49\pm0.14$	$0.12\pm0.04$	8.22	1.64
06-L-02	06.0603.07.	$998 \pm 104$	$0.63\pm0.16$	$0.08\pm0.02$	7.54	1.51
07-G-02	20.0604.07.	$360\pm46$	$0.48\pm0.06$	$0.09\pm0.02$	7.71	1.54

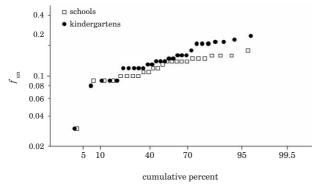


Fig. 3. Lognormal plots of the unattached fraction of radon decay products  $(f_{un})$  in Slovenian schools and kindergartens.

te-out of aerosol and, consequently, lowering F. As expected<sup>14,15</sup>, high F values are accompanied by low  $f_{\rm un}$  values. The fact that  $f_{\rm un}$  values in schools were lower than in kindergartens was therefore expected (Table 1).

Following the ICRP-65 criteria<sup>16</sup>, the personnel in kindergartens and schools can be considered as nasal breathers and therefore for them  $DCF_D = DCF_{Dn}$ . If the arithmetic means of  $f_{un}$  for kindergartens and schools (Table 1) are used in Eq. 2, values of 8.72 mSv WLM<sup>-1</sup> and 8.38 mSv WLM<sup>-1</sup>, respectively, are obtained, being higher by factors of 1.74 and 1.68, respectively, than the ICRP recommended value of 5 mSv WLM<sup>-1</sup> for work-places.

In the Postojna Cave, the patterns (not shown here) of temporal variations in summer differ substantially from those in winter, with  $C_{\text{Rn}}$  and  $C_{\text{RnDP}}$  values being lower in winter than in summer. This is because of the so called

chimney effect: in winter, the temperature outdoor is lower than in the cave, thus causing a natural draught of air from the cave through vertical channels into the outdoor atmosphere. In summer, the situation is reverse and this draught is minimal, if any. The opposite is true with F, which is lower in summer and higher in winter. The number of visitors is much higher in summer than in winter, thus causing a higher plate-out of RnDP and reducing F. Walking visitors cause more air movement in the narrow corridor at the lowest point than in the big hall at the railway station. Therefore, plate-out of RnDP is enhanced and F reduced, the effect being more pronounced in summer than in winter, and more at the lowest point than at the railway station.  $f_{un}$  values are higher in summer than in winter at the lowest point, and higher at the lowest point than at the railway station in summer. Because of the much larger number of visitors in summer than in winter, the cave air is much more disturbed in summer, resulting in enhanced  $f_{un}$  values in summer<sup>17</sup>. Similarly, the air is more disturbed by visitors in narrow corridors at the lowest point than in the big hall at the railway station, resulting in enhanced  $f_{\rm un}$  at the lowest point<sup>17</sup>.  $f_{un}$  values in the cave are much higher than in kindergartens and schools. One of the reasons for the high values is the very low concentration of aerosols in Postojna Cave, being only  $(1.9-4.3) \times 10^9$  m<sup>-3</sup>, as compared to a barite mine with  $(200-1600) \times 10^9$  m<sup>-3</sup> and  $f_{\rm un}$ in the range from 0.003 to 0.008<sup>1</sup>. Another reason for elevated  $f_{un}$  values could be the intensive washout of aerosols at practically 100 % relative humidity of the cave air, as observed in some spas<sup>18</sup>.

Based on the arithmetic mean values of  $f_{un}$ ,  $DCF_{Dm}$  (Eq. 1) and  $DCF_{Dn}$  (Eq. 2) values were calculated for the

Season, year (site)	$f_{ m un}$	$DCF_{\rm Dm}$ / mSv WLM <sup>-1</sup>	$DCF_{ m Dm}/5$	$DCF_{\text{Dn}} / \text{mSv WLM}^{-1}$	$DCF_{ m Dn}/5$
Summer, 1998 (LP)	0.58	61.4	12.3	15.9	3.2
Winter, 1998 (LP)	0.10	16.1	3.2	7.9	1.6
Summer, 1999 (LP)	0.60	63.3	12.7	16.3	3.3
Winter, 1999 (LP)	0.14	19.9	4.0	8.6	1.7
Summer, 2001 (LP)	0.68	71.5	14.3	18.2	3.6
Summer, 2001 (RS)	0.15	20.9	4.2	8.7	1.7

lowest point in summer and winter, and in summer at the lowest point and railway station (Table 4). According to the ICRP criteria<sup>16</sup> only maintenance workers engaged in hard physical work in the cave may be considered as mouth breathers, and for them  $DCF_{\rm D} = DCF_{\rm Dm}$ , while for all others (i. e., tourist guides, souvenir vendors, locomotive drivers)  $DCF_{\rm D} = DCF_{\rm Dn}$ . Table 4 also displays factors by which  $DCF_{\rm Dm}$  and  $DCF_{\rm Dn}$  values exceed the ICRP recommended value of 5 mSv WLM<sup>-1</sup> for workplaces.

The values of  $f_{\rm un}$  in wineries are similar to those in kindergartens and schools, or even lower, although a reliable comparison is not possible because of the small number of measurements in wineries. Values are much lower than in the Postojna Cave – obviously because of higher aerosol concentrations and lower air humidity in a winery than in the cave<sup>1,18</sup>. In contrast to other workplaces, rooms in wineries are not attended by workers every day and therefore typical diurnal variations of the measured parameters, reflecting working regime, have not been observed. According to the ICRP criteria<sup>16</sup>, workers in wineries may be considered as nasal breathers and for them  $DCF_{\rm D} = DCF_{\rm Dn}$ , calculated from Eq. 2.  $DCF_{\rm Dn}$  values in Table 3 are 1.5-to 1.9-fold higher than 5 mSv WLM<sup>-1</sup>.

## Conclusion

Measurements of  $f_{un}$  in Slovenian kindergartens have shown that  $DCF_D$  is on average 8.72 mSv WLM<sup>-1</sup>, which is 1.74-fold higher than the epidemiology-based value of

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 $5 \text{ mSv WLM}^{-1}$  currently recommended by ICRP for workplaces. For schools this factor was 1.68.

In the Postojna Cave,  $DCF_D$  values were higher than  $DCF_E$  by a factor of 1.7 in summer at the railway station, and by a factor from 3.2 to 3.6 in summer and from 1.6 to 1.7 in winter at the lowest point. For maintenance workers engaged in hard physical work this factor is 4.2 in summer at the railway station, and from 12.3 to 14.3 in summer and from 3.2 to 4.0 in winter at the lowest point.

In the four wineries,  $DCF_{\rm D}$  values were in the range from 7.71 mSv WLM<sup>-1</sup> to 9.56 mSv WLM<sup>-1</sup> and thus between 1.5- and 1.9-fold higher than  $DCF_{\rm E}$ .

The  $DCF_{\rm D}$  values in the different environments contribute additional information to the general database on  $DCF_{\rm D}$ , which is a prerequisite for a better understanding of and successful coping with the gap between  $DCF_{\rm D}$  and  $DCF_{\rm E}$ , which is not yet completely understood.

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# RAZINE PRODUKATA RASPADANJA RADONA NANOVELIČINE U ZRAKU INTERIJERA: USPOREDBA KOD RAZLIČITIH OKOLIŠA

# SAŽETAK

Udio nevezanih ( $f_{un}$  – unattached fraction) kratkoživućih produkata raspadanja <sup>222</sup>Rn nanoveličine, mjeren je u zraku različitih interijera te su dobiveni slijedeći rasponi: 0,03–0,25 u dječjim vrtićima, 0,03–0,18 u školama, 0,08–0,20 u vinarijama, i 0,12–0,68 u krškim špiljama. Istraživana je i raspravlja se ovisnost  $f_{un}$  o okolišnim uvjetima (temperatura zraka, relativna vlažnost i tlak), kao i o radnom režimu. Primjenjujući pristup dozimetrije, izračunati su faktori pretvorbe doze na temelju izmjerenih vrijednosti  $f_{un}$  i uspoređeni s vrijednostima dobivenima u epidemiološkim istraživanjima, koji se trenutačno preporučuju za uporabu u dozimetriji radona.